

Original Research Article

Distribution of Polycyclic Aromatic Hydrocarbons and Heavy Metals in Soils from Municipal Solid Waste Land Fill

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ABSTRACT

Aims: The distribution of polycyclic aromatic hydrocarbons and heavy metals in soils from Onitsha municipal solid waste land fill were studied.

Study design: The Onitsha solid waste dumpsite were divided into two groups, the active site labelled (A) where waste dumping is still actively practised and the dormant site labelled (B) with passive or withdrawing waste dumping. Soil sample were obtained in triplicate at each site at depth of 0-15 cm using calibrated soil auger. At each sampling site, the surface debris was removed and subsurface soil dug to 0-15 cm. Then 20 grams of soil samples from each depth was taken into a sterile container and transported to the laboratory for analysis.

Place and Duration of Study: Analysis was done at the Department of Biochemistry, Anambra, State University, Department of Applied Biochemistry, Nnamdi Azikiwe University, Awka and Spring Board Research Laboratory, Awka between February 2014 and March 2014.

Methodology: Polycyclic Aromatic Hydrocarbons (PAHs) were determined simultaneously using high resolution gas chromatography/mass spectrometry (GC/MS) (GC/MS Shimadzzi QP 2010). Heavy metals were analysed using Atomic Absorption Spectrophotometer (AAS).

Results: The result showed that Low molecular weight PAHs were not detected in both soils of active and dormant dumpsite. High molecular weight PAHs (HMW PAHs) were predominant in soil from active site among the HMW PAHs, Benzo(a)pyrene was highest followed by Benzo(k)fluoranthene and Indeno(1,2,3-cd)pyrene and the least was 1,2 Benzo(a)anthracene. The medium molecular weight PAHs (MMW PAHs) Fluoranthene occurred sparingly in the active dumpsite. The result of heavy metal concentration in the soil were in the order of Fe > Cu > Zn > Cr > Pb > Ni > Cd.

Conclusion: PAHs and Heavy metal concentration were found to be very high and as such the practise of cultivating vegetables on soil from these dumpsites for agricultural purposes should be discouraged.

Keywords: PAHs, heavy metals, soil pollution, solid waste dump, toxicant.

1. INTRODUCTION

Nearly all human activities generate waste and the way in which this is handled, stored, collected and disposed of can pose threat to the environment and to public health [1]. Waste land filling open dump is the most common method of disposal of municipal solid waste (MSW) in Nigeria. This is a practice adopted as a substitute to ocean out fall of sewage

domestic and industrial waste due to the effect of the latter on lives in the ocean [2]. Several fluxes of waste from different series and up at the open dumpsite contain a variety of contaminants which can pollute the soil of the area [3]. Soil is an environmental medium which is subjected to a number of pollutants due to different human activities [4]. Soil is one of the repositories for anthropogenic waste. Biochemical processes can mobilize the chemical substances contained in it to pollute water supplies and impact food chains among persistent organic pollutant in soil, polycyclic aromatic hydrocarbons (PAHs) and heavy metals contamination in the soils is a major concern because of their toxicity and threat to human life and the environment.

Onitsha, the largest and most populous city in Anambra State is the commercial nerve centre of Anambra State [5]. It has several manufacturing businesses like, plastics, canned bottled drinks, pharmaceutical products, batteries, electrical wire and cables, paints, greases, motor oils and petroleum products. It is a home of different auto spare parts both old and new, electrical and electronic products, building materials and hundreds of petrol stations. In addition, a huge artisan population exists in the city to render welding, soldering, brazing, sheet metal working, electrical and other services to the public. Also used consumer products such as electronics, light bulbs, aerosols, cans, lead foils, bottles, caps garbages and other materials are continually generated on a large scale by residents of the city. All these contribute significant quantities of PAHs and heavy metals to the waste of the city. Polycyclic aromatic hydrocarbons (PAHs) (Table 1) are one of the most common classes of organic pollutants detected in open dumpsite [6, 7]. They are introduced into the landfill along with petroleum hydrocarbon residues [8] and by deposition from the atmosphere [9], but they can also be formed on the spot in municipal landfill fires [10]. They can occur in the air either attached to dust particles or as solids in soil or sediments [11]. The occurrence of various heavy metals such as As, Cd, Pb, Ni, Zn, Fe, Cu, Cr etc in MSW are of concern and was reported by many workers [12-14].

Since these contaminants affect the environmental qualities in addition to affecting plant and animal health in and around open dump sites, monitoring of soil qualities especially PAHs and heavy metal distribution in dump site becomes necessary in order to facilitate recommendation of suitable remedial measures [15, 16]. There are very few reports on PAHs and heavy metal status in soil from active and dormant solid waste dump site in Anambra State. These studies aimed at determining the distribution of PAHs and heavy metals in soils from MSW dump. A significant importance of this work will be in providing baseline information for further soil quality monitoring studies.

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Table 1. Classification of Polycyclic Aromatic Hydrocarbon (PAHs) based on Molecular Weight [17]

PAHs	Chemical Formular	Molecular Weight	Number of Rings
Low Molecular Weight (LMW)			
Naphthalene	C ₁₀ H ₈	128.00	2
Acenaphthylene	C ₁₂ H ₈	152.00	3
Acenaphthene	C ₁₂ H ₁₀	154.21	3
Fluorene	C ₁₃ H ₁₀	166.20	3
Phenanthrene	C ₁₄ H ₁₀	178.20	3
Anthracene	C ₁₄ H ₁₀	178.20	3
Medium Molecular Weight (MMW)			
Fluoranthene	C ₁₆ H ₁₀	202.26	4
Pyrene	C ₁₆ H ₁₀	202.30	4
High Molecular Weight (HMW)			
Chrysene	C ₁₈ H ₁₂	228.30	4
Benzo(a)anthracene	C ₁₈ H ₁₂	228.89	4
Benzo(k)fluoranthene	C ₂₀ H ₁₂	251.30	5
Benzo(k)fluoranthene	C ₂₀ H ₁₂	251.30	5
Benzo(a)pyrene	C ₂₀ H ₁₂	251.30	5
Indono(1,2,3-Cd)pyrene	C ₂₂ H ₁₂	276.30	6
Benzo(g,h,i)pevykene	C ₂₂ H ₁₂	276.30	6
Dibenzo(a,h)anthracene	C ₂₂ H ₁₂	278.35	6

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61 2. MATERIAL AND METHODS

62 2.1 Site Description

63 The present study focused on Onitsha municipal solid waste open dump site located at lat
64 06° 6' N and long 06° 48' E opposite National Metallurgical Institute along Onitsha-Owerri
65 road Onitsha, Anambra State, Nigeria (Figure 1). The dumpsite is surrounded by factories;
66 gasoline stations, motor parks, auto spare parts market, auto repair workshops, electrical/
67 electronic market, hospitals and house.

68 The dumpsite (Figure 2) contains mixtures of both organic and inorganic waste materials,
69 such as food wastes, papers, cardboards, metals, tins, glasses, ceramics, battery wastes,
70 textile rags, plastics, sewage might soils and other miscellaneous materials such as bricks,
71 ash, fine dust, rubber and wood waste etc.

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Figure 1: Map of Anambra State, Nigeria showing the study area.



Figure 2: The Municipal solid waste dump site located along Onitsha-Owerri road, Anambra State, Nigeria.

83 2.2 Field reconnaissance and sampling

84 The goal of the field reconnaissance survey was to ascertain the level of pollution on the
 85 solid waste dumpsite. Sampling was subsequently carried out by mapping out plots on the
 86 dumpsite using the grid system. The dumpsite were divided into two groups, the active site
 87 labelled (A) where waste dumping is still actively practised and the dormant site labelled (B)
 88 with passive or withdrawing waste dumping. Soil sample were obtained in triplicate at each
 89 site at depth of 0-15cm using calibrated soil auger. At each sampling site, the surface debris
 90 was removed and subsurface soil dug to 0-15 cm. Then 20 grams of soil samples from each
 91 depth was taken a sterile container and labelled. The entire sample was transported to the
 92 laboratory for analysis.

94 2.3 Polycyclic Aromatic Hydrocarbons (PAHs) Analysis

95 Sixteen non-substituted PAHs (ns, PAHs) with two to five aromatic rings Naphthalene,
 96 Acenaphthylene, Acenaphthene, Fluorene, Phenanthrene, Anthracene, Fluoranthene,
 97 Pyrene, Chrysene, 1,2-benzoanthracene, Benzo(b)fluoranthene, Benzo(k)fluoranthene,
 98 Benzo(a)pyrene, Indeno(1,2,3-cd)pyrene, 1,25-Dibenzoanthracene, 1,2-Benzo(a)perylene
 99 were determined simultaneously using high resolution gas chromatography/mass
 100 spectrometry (GC/MS) as reported by Tang et al. [16].

101 Before GC/MS analyses, 10g of soil samples were extracted with 50 ml extraction solvent
 102 mix of acetone and dichloromethane (50:50 v/v) using an ultra-sonic bath and the extraction
 103 was subsequently repeated with 50 ml of the extraction solvent. The sample extract was
 104 concentrated using rotary evaporator and then reconstituted with Cyclohexane and
 105 reconcentrated to 2 ml and transferred into Cyclohexane and purified using a Silica gel
 106 column. The final fractions containing PAHs was eluted with 25 ml of dichloromethane
 107 pentane mix (2:3 v/v) and analyzed GC/MS (Shimadzzi QP 2010) equipped with a 30 m x
 108 0.25 mm ID fused silica capillary column coated with 0.25 μ m thick film of HP-5.

109 2.4 Heavy metal Analysis

110 Heavy metals were determined using Atomic Absorption Spectrophotometer as described by
 111 Adelekan and Alawode [14]. The fresh soil was spread on a clean plastic sheet placed on a
 112 flat surface and air dried in open air in the laboratory under room temperature for 24 hrs.
 113 Afterwards the soil was sieved using 90.5 mm sieve. After sieving, 0.5g was weighed from
 114 the sieved soil of each sample into a Pyrex 250ml council flask. Ten (10) ml of Perchloric
 115 and Nitric acid mixture in ratio 1:2 was added to the sample and was digested on a hot plate
 116 at 105 °C for 30 minutes in a fume cupboard. After digestion the flask was left to cool and
 117 the digest was transfered into a 50 ml vial and made up to 50 ml mark with deionized water.
 118 The resulting solutions were then analysed using Atomic Absorption Spectrophotometer
 119 AAS (GBC Avanta ver 2.02).

120 2.5 Statistical Analysis

121 Data obtained were expressed as mean $\pm$ standard deviation (SD) and One Way Analysis of
 122 Variance (ANOVA) was used to compare obtained means. Values with $P \leq 0.05$ were
 123 considered statistically significant.

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126 3. RESULTS AND DISCUSSION

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128 3.1 Results

129 The analysis of the heavy metal concentrations of the Active and Dormant dumpsite is
 130 presented in Table 2. The heavy metal concentrations of the Active dumpsite are as follows
 131 Fe (7460.00±10.00), Zn (362.75±0.43), Cu (1976.50±0.50), Cr (94.67±0.15), Pb
 132 (50.00±0.50), Cd (5.90±0.10), and Ni (30.10±0.10). The heavy metal concentrations of the
 133 dormant site are Fe (457.18±0.50), Zn (214.43±0.51), Cu (322.47±0.50), Cr (29.41±0.52),
 134 Pb (28.42±0.52), Cd (1.85±0.30), and Ni (11.10±0.09).

135 **Table 2. Heavy Metal Concentration (mg/kg) in Soil from active and dormant dumpsite**

Heavy Metals	Active site	Dormant site
Fe	7460.00±10.00 ^a	457.18±0.50 ^b
Zn	362.75±0.43 ^a	214.43±0.51 ^b
Cu	1976.50±0.50 ^a	322.47±0.50 ^b
Cr	94.67±0.15 ^a	29.41±0.52 ^b
Pb	50.00±0.50 ^a	28.42±0.52 ^b
Cd	5.90±0.10 ^a	1.85±0.30 ^b
Ni	30.10±0.10 ^a	11.10±0.09 ^b

136 **Values are mean±standard deviation of triplicate determination. Values in the same row*
 137 *bearing the same superscript letter are not significantly different at p< 0.05.*

138 The result of the PAHs analysis is presented in Table 3.

139 **Table 3. Heavy PAHs Concentration (mg/kg) in Soil from active and dormant dumpsite**

140

PAHs Constituents	Active Site	Dormant Site
Low Molecular Weight		
Naphthalene	0.02±0.01	ND
Acenaphthylene	ND	ND
Acenaphthene	ND	ND
Fluorene	ND	ND
Phenanthrene	ND	ND
Anthracene	ND	ND
<i>Sub total low molecular weight</i>	<i>0.02±0.01</i>	<i>ND</i>
Medium Molecular Weight		
Fluoranthene	0.02±0.02	ND
Pyrene	ND	ND
<i>Sub total</i>	<i>0.02±0.02</i>	<i>ND</i>
High Molecular Weight		
Chrysene	0.04±0.02	0.01±0.01
1,2 Benzo(a)anthracene	0.02±0.00	ND
Benzo(b) fluorene	0.03±0.02	0.01±0.00
Benzo(c) fluoranthene	0.05±0.01	0.03±0.02
Benzo(a)pyrene	0.11±0.01	0.04±0.02
Indeno(1,2,3CD)pyren	0.05±0.01	0.02±0.01
1,2,5 Dibenzoanthracene	0.04±0.01	0.01±0.01
1.2 Benzo(a)perylene	0.04±0.02	0.01±0.00

141 **Values are mean±standard deviation of triplicate determination. ND = Not Detected.*

142 **Table 4. Correlation of HMW PAHs and heavy metal concentration in soil from MSW**
143 **dumpsite**

	Fe	Zn	Cu	Cr	Pb	Cd	Ni	Chry	Baa	Bbf	Bkf	Bap	Indp	Dah	Bapy
Fe	1														
Zn	-.866	1													
Cu	-.500	.000	1												
Cr	-.655	-.1189	-.982	1											
Pb	-.500	.000	.1.000*	-.982	1										
Cd	-.500	.000	- .000*	-.982	1.000*	1									
Ni	.500	.000	-1.000*	.982	-1.000*	-1.000*	1								
Chry	-.500	.000	1.000*	-.982	1.000*	1.000*	-1.000*	1							
Baa	.000	.500	-.866	.756	-.866	-.866	.866	-.866	1						
Bbf	.000	.500	-.866	.756	-.866	-.866	.866	-.866	-1.000*	1					
Bkf	-.500	.000	1.000*	-.982	1.000*	1.000*	-1.000*	1.000*	-.866	-.866	1				
Bap	-.655	.189	.982	-1.000*	.982	.982	-.982	.982	-.756	-.756	.982	1			
Indp	.000	.500	-.866	.756	-.866	-.866	.866	-.866	-1.000*	1.000*	-.866	-.756	1		
Dah	.000	.500	-.866	.756	-.866	-.866	-.866	.866	1.000*	1.000*	-.866	-.756	1.000*	1	
Bapy	-.500	.000	1.000*	-.982	-1.000*	1.000*	1.000*	-1.000*	-.866	-.866	-1.000*	.982	-.866	-.866	1

144 *Chry = Chrysene, Bbf = Benzo(b)fluoranthene, Bkf = Benzo(k)fluoranthene, Bapy =
145 Benzo(a)pyrene, Indp = Indeno(1,2,3-cd)pyrene, Dah = 1,25-Dibenzoanthracene, Bap = 1,2-
146 Benzo(a)pevydene. (* = correlation is significant at the 0.01 level)..

147 3.2 Discussion

148 The result of heavy metal concentration in soil from both active and dormant site of MSW
149 dump indicated that heavy metal concentration in active site were significantly higher at ($p <$
150 0.05) than heavy metal concentration in soil from dormant site (Table 2). The heavy metal
151 range were on the order of $Fe > Cu > Zn > Cr > Pb > Ni > Cd$. The differences in
152 concentration of heavy metals in both sites may be attributed to difference in age of the
153 dump site. The active site still receives waste on daily basis while waste dumping has been
154 suspended at the dormant site for over five years. Other factors include plant uptake, soil
155 pH, organic matter content and other soil characteristics affects the amount of heavy metal
156 leaching [14].

157 Both dumpsites had the highest Fe concentration (475.18 ± 0.74 - 7460.00 ± 10.00 mg/kg).
158 This value was higher than 15.52 - 103.46 mg/kg as reported by Tripathi and Misra [18] in
159 dump site at Allahabad India and the value 56.12 - 280.16 mg/kg reported by David et al.
160 [19] in dump site in Part Harcourt, Rivers State, Nigeria. Concentration of Zn in soil ranged
161 from 214.43 ± 0.51 - 362.75 ± 0.43 mg/kg which was higher than the values 18.21 - 108.00
162 mg/kg reported by Tripathi and Misra [18] and 21.18 - 128.16 mg/kg reported by David et al.
163 [19]. Cu values ranged from 322.47 ± 0.50 - 1979.50 ± 0.50 mg/kg in this study and were
164 higher than values reported in literature by Tripathi and Misra [18] and David et al. [19].
165 Elevated levels of these metals in soil from the dump site may be attributed to deposition of
166 solid waste, incidental accumulation, atmospheric depositions and deposition of agricultural
167 chemicals [20].

168 Fe, Zn and Cu are essential trace metals required by the body for metabolic activities. Fe
169 and Zn are cofactors to some metabolic enzymes in both plants and animals. In case of
170 women, pregnancy and lactation increases the demand for Fe, Zn and Cu [21, 22].
171 Decrease in the store of these essential trace metals may aggravate Cd and Pb toxicity [23].
172 High levels of zinc above maximum Zn tolerance for human health could produce zinc
173 toxicity which could result to low copper status, altered ion function and reduced immune
174 function while high levels of Cu and Fe may result to nausea, vomiting, diarrhea and
175 abdominal pains [24, 25].

176 Concentrations of Cr in soil from active dump site (94.67 ± 0.15 mg/kg) were significantly
177 higher ($P < 0.05$) than the value from the dormant site (29.41 ± 0.52 mg/kg). The values at the
178 active site were higher than the values 6.25 - 75.55 mg/kg reported by Adelekan and
179 Alawole [14] in dumpsite within Ibadan in Southwest Nigeria but lower than the values 900 –
180 2000 mg/kg reported by Adefemi and Awokunmi [26]. The values of Cr obtained from this
181 study were higher than Maximum Allowable Limits (MAL) of Cr concentration 60 mg/kg in
182 Germany and Sweden and 30 mg/kg in Denmark and Netherlands [27]. The high
183 concentration of Cr in active dumpsite may be attributed to increase in industrial and
184 municipal waste which contained high concentration of Cr. Accumulation of Cr in the body
185 can results to human health risk.

186 The value of Pb from the study ranged from 28.42 ± 0.52 – 50.00 ± 0.50 mg/kg. The values of
187 Pb from the active dump site was high compared to the MAL of 40.00 mg/kg set for countries
188 like Denmark, Netherland and Sweden [27] but lower than the limit 70 mg/kg, 100 mg/kg set
189 for Germany and France respectively [27]. The increase in Pb in the active dumpsite may be
190 due to atmospheric deposition owing to its very close surrounding of high vehicular traffic
191 highway and dumping of waste with high Pb content. Environmental protection Agency
192 (EPA) [28] reported lead as a probable human carcinogen. Elevated levels of Pb in human
193 reduces plasma copper which may lead to irreversible damage of the brain, alter the

flexibility of red blood cells (RBC), increase cell fragility leading to increased risk of haemolysis [24, 29].

Concentrations of Cd in soil obtained from active site were significant higher at $P < 0.05$ than the dormant site. This may be attributed to differences in the age of the dumpsite and biodegradation of the dormant dumpsite. The value of Cd from the study ranged from 1.85 ± 0.30 - 5.90 ± 0.30 mg/kg. These values were higher than the MAL for Cd 1.00 mg/kg in Spain and Germany and 0.5 mg/kg in Denmark and Netherland [27]. Indiscriminate dispose of Cd containing waste on the dumpsite and atmospheric deposition may contribute to the high level of Cd in the dumpsite. Accumulation of Cd in humans may lead to acute cadmin poisoning which may cause high blood pressure, kidney damage, destruction of testicular tissues and destruction of red blood cells [30]. Adelekan and Alawole [14] reported that cadmium, lead and mercury can be harmful to animals and humans at relatively low concentrations and thus should receive close scrutiny in relation to application of municipal solid waste composts to Agricultural soil. The concentration of Ni in the dumpsite ranged from 11.10 ± 0.09 to 30.10 ± 0.10 mg/kg.

Concentration of Ni were found to be significantly highest ($P < 0.05$) in soil from active dumpsite than in soil from dormant dumpsite. The values from the active site (30.10 ± 0.10 mg/kg) were higher than the values (5.65-19.40 mg/kg) reported by Adelekan and Alawole [14] in soil from dumpsite in Ibadan. The value of Ni found in the active site was higher than the MAL of 15 mg/kg found in countries like Denmark and Netherland but within MAL of 50 – 70 mg/kg as in Austria [27] and 75 mg/kg as in Australia and New Zealand [31]. Elevated level of Ni in soil from active dumpsite above that of dormant site could be as a result of continuous dumping of Ni compounds and alloys at the active dumpsite and biodegradation as result of aging of the dormant site. Accumulation of Ni in human through inhalation, ingestion or skin absorption may result to cancer of the lung larynx and prostrate. Others include dermatitis, bronchial asthma, myocardial infarction, respiratory failures and birth defects [29, 31].

Result of polycyclic aromatic hydrocarbon (PAHs) distribution in the MSW dumpsites are represented in (Table 3). Low molecular weight PAHs (LAW PAHs) were not detected in both soils of active and dormant dumpsite. High molecular weight PAHs (HMW PAHs) were predominant in soil from active site along the HMW PAHs, Benzo(a)pyrene was highest followed by Benzo(k)fluoranthene and Indeno(1,2,3-cd)pyrene and the least was 1,2 Benzo(a)anthracene. The medium molecular weight PAHs (MMW PAHs) Floranthene occurred sparingly in the active dumpsite. The presence of HMW PAHs compared to LMW PAHs may be due to the recalcitrant nature of the HMW PAHs, age of dumpsite, weather and climatic condition. Low molecular weight PAHs such as Naphthalene, Acenaphthylene Acenaphthene, Fluorene, Anthracene and Phananthrene are transformed rapidly by many bacteria and fungi while high molecular weight PAHs however are more recalcitrant in the environment and resist both chemical and microbial biodegradation [33].

The predominance of HMW PAHs in the active site compared to dormant site may be attributed to differences in biological diversity, abundance and the age of pollution given that the dormant site was not receiving waste. Therefore PAH constituents must have volatilized and disappeared from the soil but the high molecular weight constituents still abound but in much lower concentration than in active dumpsite. It is well known that light PAHs compound occur in soil in lower proportions than heavier ones due to their physicochemical properties such as high solubility, volatility, biodegradability and lower sorption ability to soil organic portion [34, 35]. This also corroborates the report of Reid et al. [36]; Semple et al. [37] and Dick et al. [38] which stated that the fate of PAHs in the soil is affected by factors such as biological diversity and abundance, the nature of the soil minerals and organic matter and

pollutant properties such as solubility, polarity, hydrophobicity, lipophilicity and molecular structure. This might explain the differences in PAHs concentration between soil samples that are not in close proximity as well as variation in PAHs concentration of the same site.

PAHs adsorb to organic matters, percolate the soil and contaminate the aquifers. They are taken up by aquatic species and plants and passed through food chain and bioaccumulate in human on long term exposure. This may lead to human health disorder such as organ dysfunction, reproductive, neurobehavioral and genetic disorders [39]. The international agency for research on cancer (IARC) and Environmental protection Agency (EPA) has determined the following Benzo(a)anthracene, Benzo(a)pyrene, Benzo(b)fluoranthene, Benzo(k)fluoranthene, Indeno(1,2,3-cd)pyrene as possible human carcinogen [40]. The correlation coefficient was also analysed to understand the relationship between HMW PAHs and heavy metal concentration. Various notable significant correlations were recorded and matrix of correlation coefficient between different quantitative variables is presented in (Table 4). The result of correlation can prove useful in understanding the relationship of metal concentration with each other and with PAHs distribution in the soil. The correlation study of high molecular weight PAHs and heavy metals from the study sites revealed that there is a significant positive correlation among the heavy metals Pb, Cd, and Cu and among the HMW PAHs Bbf, Chy, Ind and Dah at $P < 0.01$ level indicating that a positive linear relationship existed among the heavy metals and among the HMW PAHs suggesting that the source of the heavy metals and HMW PAHs may be the same and could be attributed to waste dump due to human activities. Also there are strong positive relationship between HMW PAHs (Bbf, Baa, Bkf, Chry Indp and Dah) and heavy metals (Pb, Cd, Cu) at $P < 0.01$ level. This showed that the HMW PAHs increased with increase in heavy metal level. The observation that significant negative correlation exist between Bap and Cr indicated that Benzo(a)pyrene increased with decrease in Chromium level. Benzo(a)pyrene one of the most recalcitrant PAHs in soil adsorbs to the soil matrix and thus is physically unavailable to degradative bacteria and fungi. Carl et al. [33] reported that the formation of non extractable bound residues is a significant sink of PAHs in soils.

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272 4. CONCLUSION

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The absence of proper management practice and lack of adequate capacity for final deposition of solid waste has been a serious problem in urban areas. The recommendation of the study therefore includes the prevention of further waste dumping at the dumpsite and relocation of dumpsites out of the city. Also proper monitoring and remediation plans should be put in place to reduce the chances of ground water pollution by leaching of these contaminants and the practise of cultivating vegetables on these dumpsite or the introduction of municipal solid waste to agricultural soil should be discouraged.

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282 REFERENCES

- 283 1. Zhu D, Asnani, PU, Zurbruggi C, Anapolsky S, Mani S. Improving Solid Waste
284 management in India. A source book for policy makers and practitioners. The World
285 Bank. Washington. DC. 2008.
- 286 2. Adriano DC. Trace elements in the terrestrial environment. 2nd ed. New York:
287 Springer Verlag. 1986; 533.
- 288 3. Sukop H, Blume HP, Kunick, W. The soil, flora and vegetation of Berlin's waste land
289 Nature in cities. In Laurie IC, editor. The natural environment in the design and
290 development of urban green space. Chichester: John Wiley and Sons. 1979; 115-
291 132.

- 292 4. Al-Khashman OA, Shawabkeh RA. Metal distribution in soils around the cement
293 factory in southern Jordan. *Environ. Pollut.* 2006; 140(3):387-394.
- 294 5. National Population Commission (NPC). Federal Republic of Nigeria official gazette.
295 2009; 96: 2.
- 296 6. Gade B, Layh M, Westermann H, Amsoneti N. Determination of organic perimeters
297 in waste and leachates from the hazardous waste landfill of Rainforf, Germany.
298 *Waste Manage. Res.* 1996; 14:553-569.
- 299 7. Kerndorff, H. Goundwater contamination assessment by problem specific selection
300 of analytical parameters. *Inter. J. Environ. and Chem.* 1995; 60:239-256.
- 301 8. Loehr, RC, Erickson DC, Kelmar LA. Characteristics of residues at hazardous waste
302 land treatment units wat. *Res.* 1993; 27:1127-1138.
- 303 9. Wild SR, Jones KC. Polynuclear aromatic hydrocarbon on the United Kingdom
304 environment – a preliminary source inventory and budget. *Environ pollut.* 1995;
305 88:91-108.
- 306 10. Ruokojani P, Ruuskanen J, Ettala M, Rahkonen P, Jarhanen J. Formation of
307 polyaromatic hydrocarbons and phychlonated organic compounds in municipal
308 waste landfill fires chemosphere. 1995; 31:3899-3908.
- 309 11. Lunch, A. The carcinogenic Effects of polycyclic aromatic hydrocarbons. London:
310 Imperial collage press. 2005; 117-125.
- 311 12. Ogundiran MB, Osibanjo O. Heavy metal concentrations in soils and accumulation in
312 plants growing in a deserted slag dumpsite in Nigeria. *African Journal of*
313 *Biotechnology* 2008; 7(17):3053-3060.
- 314 13. Amusan AA, Ige DV, Olawale, R. characteristics of soil and crop uptake of metals in
315 municipal waste dumpsites in Nigeria. *J. Hum. Ecol. Kamla Rja.* 2005; 1(2):167–171.
- 316 14. Adelekan BA, Alawode, AO. Contribution of municipal refuse dumps to heavy metals
317 concent trations in soil profile and groundwater in Ibadan Nigeria. *Journal of App.*
318 *Biosci.* 2011; 40:2727-2737.
- 319 15. Biswas AK, Kumar S, Babu SS, Bhattacharya JK, Chakra BT. Studies on
320 environmental quality in and around municipal solid waste dumpsite. *Resource*
321 *conservation and Recycling.* 2010; 55:129-134.
- 322 16. Tang LI, Tang XX, Zhu YG, Zheng, MH. Miao QL. Contamination of polycyclic
323 aromatic hydrocarbons (PAHs) in urban soils in Begying China. *Environ. Int.* 2005;
324 (6): 822-828.
- 325 17. ATSDR (Agency for toxic substances and disease registry). Atlanta, GA, 1995; 458.
- 326 18. Tripathi A, Misra DR. A study of physic-chemical properties and heavy metals in
327 contaminated soils of municipal waste dumpsites at Allahabad, India *International.*
328 *Journal of Environmental Sciences* 2012; 2(4):
- 329 19. David NO, Benjamin LH, Patrick OY. Some physicochemical and Heavy metal levels
330 in soils of waste dumpsite in Port Harcourt municipality and Environs *J. Appl. Sci*
331 *Environ. Manage.* 2009; 13(4):65-70.
- 332 20. Ferguson C, Kasamas H. Chemical properties, Health and Environmental Effects of
333 Cobalt lenntech water treatment and purification Holding B.V. 1999.
334 Available: www.lenntech.com/periodic/elements/co.htm.
- 335 21. Picciano MF. Pregnancy and lactation. In Ziegler EE, Filer LJ Jr. eds. *Present*
336 *Knowledge in Nutrition* 7th ed. Washington: DC ILSI press. 1996; 384–395.
- 337 22. National Academy of Sciences/Institute of Medicine (NAS/10M). *Dietary References*
338 *intakes for vitamin A, vitamin K, Arsenic, Boron, Chromine, Copper, Iodine, Iron,*
339 *Manganese, Molybderum, Nickel, Silicon, Vanadium and Zinc, food and Nutrition*
340 *Board, institute of medicine Washington DC. ISBN0-309-7279-4.*
- 341 23. Berglund M, Askesson A, Nevmell, B., Vahter, M. Intestinal Absorption of Dietary
342 Cadminm in women depends on Body stores and fiber intake *Environ. Health*
343 *Perspect.* 1994; 102:1058-1065.

- 344 24. Khurshid JS and Iqbal HO. The role of inorganic elements in human body Nucleus.
345 1984; 21:3-23.
- 346 25. Macrae R, Robinson RK, Sadler MJ. Encyclopdia of food science, food Technology
347 and Nutrition. New York: Academic press. 1993; 2573-2672.
- 348 26. Adefemi SO and Awokunmi EE. The impact of municipal solid waste disposal in Ado
349 Ekiti metropolis. Ekiti State Nigeria. African Journal of Environmental Science and
350 Technology 2009; 3(8):186-189.
- 351 27. European Commission Director General Environmental (ECDGE). Heavy metals
352 and organic compounds from wastes used as organic fertilizers final report. 2004;
353 73-74. WPA consulting Engineers Inc. Reference number TEND/AML/2001/07/20.
- 354 28. EPA (1989) US Environmental protection Agency code of Federal Regulation 40
355 CFR 302.4.
- 356 29. Sabirie M, Wendy G. Human health effect of heavy metals Environment, science
357 and Technology brief for citizens. Centre for hazardous substances Research
358 Kansas State University; 2009.
- 359 30. Bryce-Smith D. Heavy metals as contamination techniques environment. In Peter G.
360 editor. Educational techniques subject group. Chemical Society London. 1977; 21-
361 23.
- 362 31. ANZECC and NHMRC. Australian and New Zealand Guidelines for the Assessment
363 and management of contaminated sites published by Australian and New Zealand
364 Environment and conservation council (ZECC) and the New Zealand Health and
365 Medical Research Council (NHMRC) soil using computer-controlled scanning
366 electron microscope. Environ forensic. 1992; 3:131-143.
- 367 32. Lentech Water Treatment. Chemical properties. Health and Environment Effects of
368 Cobalt. Lentech Water Treatment and Purification Holding B.V. 2009.
369 Available: www.lentech.com/periodic/elements/co.htm
- 370 33. Carl EC, John BS. 7 – Bioremediation of Polycyclic Aromatic Hydrocarbons by
371 ligninolytic and non ligninolytic fungi. London: Cambridge University press. 2009;
372 136-187.
- 373 34. Weiss P, Lorbeer G, Scharf S. Regional aspects and statistical cauterization of the
374 load with semi volatile organic compounds at remote Austrian forest sites
375 chemosphere. 2000; 40:1159-1171.
- 376 35. Jensen J, Reimann C, Finne TE Ottesen RT, Arnoldussen A. PAH concentrations
377 and compositions in the top 2cm of forest soils along a 120km long transect through
378 agricultural areas forest and the city of OsloNorway. Environmental pollution. 2007;
379 14: 829-838.
- 380 36. Reid BJ, Jones KC, Sample, K.T. Bioavailability of persistent organic pollutant in
381 soils and sediments a perspective on mechanisms, consequences and assessment.
382 Environ. Pollut. 2000; 108:103-112.
- 383 37. Semple KT, Morriss AWJ, Gaskin, IP. Bioavailability of hydrophobic organic
384 contaminates in soils fundamental concepts and techniques for analysis Eur. J. Soil
385 sci. 2007; 54: 809-818.
- 386 38. Dick J, Klingelmann E, Burauel P, Jones KC, Samples KT. Long-term fate of
387 polychlorinated biphenyls and polycyclic aromatic hydrocarbons in agricultural soil
388 Environ. Sci. Technol. 2005; 39:3663-3670.
- 389 39. Palmer SR, Dunstan FDJ, Fielder H, Fone DL, Higgs, G, Senior, M.L. Environ. 2005.
- 390 40. Emily M. Health effect of polycyclic aromatic hydrocarbons in Encyclopedia of Eorth
391 cutler J. Cleveland Eds. Environmental information coalition. National council for
392 science and the Environment Washington DC. 2008.